

# Effective TiO<sub>2</sub>-Sulfonated Carbon-derived from Eichhornia crassipes in The Removal of Methylene Blue and Congo Red Dyes from Aqueous Solution

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**Submission date:** 09-Mar-2022 08:14AM (UTC+0700)

**Submission ID:** 1779844828

**File name:** BCREC\_lis\_Intan\_Widiyowati,\_Mukhamad\_Nurhadi.pdf (981.53K)

**Word count:** 8758

**Character count:** 41166



Research Article

## Effective TiO<sub>2</sub>-Sulfonated Carbon-derived from *Eichhornia crassipes* in The Removal of Methylene Blue and Congo Red Dyes from Aqueous Solution

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Received: 4<sup>th</sup> January 2020; Revised: 26<sup>th</sup> May 2020; Accepted: 27<sup>th</sup> May 2020;  
Available online: 30<sup>th</sup> July 2020; Published regularly: August 2020

### Abstract

The study of TiO<sub>2</sub>-sulfonated carbon-derived from *Eichhornia crassipes* (TiO<sub>2</sub>/SCEC), as an effective adsorbent to remove Methylene blue (MB) and Congo red (CR) dyes from aqueous solution, has been conducted. The preparation steps of TiO<sub>2</sub>/SCEC adsorbent involved the carbonisation of *E. crassipes* powder at 600 °C for 1 h, followed by sulfonation of carbon for 3 h and impregnation through titanium(IV) isopropoxide (500 μmol). The physical properties of the adsorbents were characterized by using X-ray fluorescence (XRF), Fourier transform infrared, X-ray diffraction (XRD), Scanning electron microscopy with Energy dispersive X-ray (SEM-EDX), Thermogravimetric analysis (TGA) and nitrogen adsorption-desorption studies. The dye removal study using TiO<sub>2</sub>/SCEC adsorbent was carried out by varying of contact time, adsorbent dosage, initial dye concentration, pH, particles size of adsorbent and temperature. The kinetics models were determined by the effects of contact time and the thermodynamic parameters ( $\Delta H$ ,  $\Delta S$ , and  $\Delta G$ ), which were calculated by the effects of temperature. The results showed that the maximum dye removal capacity of TiO<sub>2</sub>/SCEC were 18.8 mg.g<sup>-1</sup> for MB and 36.5 mg.g<sup>-1</sup> for CR. The removal of MB and CR dyes using TiO<sub>2</sub>/SCEC adsorbent performed a pseudo-second order kinetic models with spontaneity. Copyright © 2020 BCREC Group. All rights reserved

**Keywords:** *Eichhornia crassipes*; TiO<sub>2</sub>; methylene blue; congo red; adsorption; carbon

**How to Cite:** Widiyowati, I.I., Nurhadi, M., Hatami, M., Yuan, L.S. (2020). Effective TiO<sub>2</sub>-Sulfonated Carbon-derived from *Eichhornia crassipes* in The Removal of Methylene Blue and Congo Red Dyes from Aqueous Solution. *Bulletin of Chemical Reaction Engineering & Catalysis*, 15(2), 476-489 (doi:10.9767/bcrec.15.2.6997.476-489)

**Permalink/DOI:** <https://doi.org/10.9767/bcrec.15.2.6997.476-489>

### 1. Introduction

Synthetic dyestuff like Methylene blue (MB) and Congo red (CR) are very dangerous for human and animal life if discharged into the environment. Synthetic dyestuff seldom are re-

moved as liquid waste in large volumes from many industries [1]. Liquid waste that polluted by Methylene blue (MB) and Congo red (CR) is commonly produced in textile, paint, papers, hair colorant, leather industries. The amount of dyes in the liquid waste from industries depending on many factor such as dyeing methods, type of coloured raw material (type of fabric), as well as type and amount of dyes. The quality of water became decrease with a very small

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amount of synthetic dyestuff (ppm) dissolved in the water, as they affect the colour of water and disqualify it for home and technological purposes. A large number of synthetic dyestuff into the environment can become serious problem due to it causes infertility of the soil, increased amount of chemical oxygen demand in water, and reduced light penetrability that can inhibition the photosynthetic processes [2,3]. Some synthetic dyestuffs have negative effect for living human and organism if their presences in water due to their properties are carcinogenic and mutagenic. Many problems for human health were caused by dyes such as an adverse effect on the functioning of liver, reproductive system, and kidneys, increased heart rate, vomiting, shock, jaundice, chromosomal damage, respiratory enzymes impairment, and the formation of tumors [2,4,5].

Many different methods have been used in removing synthetic dyestuff from wastewater in order to reduce their impacts on the environment. The wastewater treatment methods for removal of pollutants from aqueous solution like photochemical degradation [6], biological degradation [7], chemical oxidation [8], electrocoagulation [9], coagulation and flocculation [10], adsorption [3,11-13], ozonation [14], fungal decolorization [15], have been investigated with varying degree of success. Among of wastewater treatment technologies, adsorption is rapidly gaining prominence as physicochemical method of treating aqueous effluent due its proven efficiency and great potential as means of producing quality effluent [16].

*Eichhornia crassipes* (water hyacinth) is a weed plant with a high growth rate that can damage the aquatic environment, such as in many tropical lake and river. Its existence can disrupt fisheries, transportation and hydroelectric power production. Several study have been conducted with *E. crassipes* raw material to produce dye adsorbent, such as: *E. crassipes* for removal of azo and anthraquinone dyes from aqueous solution [17], the root of *E. crassipes* for adsorption of Congo red dye from aqueous solution [16], *E. crassipes* for adsorption of Methylene blue dye from aqueous solution [2]. However, the results from the previous research show that the dye removal capacity is low ( $Q_m = 1.58 \text{ mg.g}^{-1}$ ) and the amount of adsorbent (1 g, 50 mL of dye solution) that used is not economical.

Here, we report the usage of carbon from *E. crassipes*, which were modified by sulfonation and  $\text{TiO}_2$  impregnation, as adsorbent for the re-

moval of Methylene blue and Congo red dyes from aqueous solution.  $\text{TiO}_2$  is one of the promising adsorbent widely adopted to adsorb dyes, such as Methylene Blue, Methyl Orange, Malachite Green, Crystal Violet, Thymol Violet, Acid Red G, Reactive Red 195, etc., owing to it is comprised of hydroxyl groups, which could induce electrostatic charge and Van der Waals forces with those dyes [18-22]. However, the hydroxyl groups might be limited for high amount dyes adsorption. Hence,  $\text{TiO}_2$  is further modified with carbon source via sulfonation to enhance the adsorption capability. The function of sulfonation process was to enhance the number of oxygen functional groups that can serve as the adsorption sites, while increasing hydrophilicity [23]. The oxygen functional groups onto sulfonated carbon from *E. crassipes* act an importable in uniformly dispersing  $\text{TiO}_2$  on the surface and tailoring the particle size of  $\text{TiO}_2$ , which result in great enhancement of adsorption capacity of dyes due to higher surface area.  $\text{TiO}_2$  was impregnated onto sulfonated carbon from *E. crassipes* due to  $\text{TiO}_2$  has a high activity to adsorb dyes pollutants in water, low cost, non-toxicity, chemical stability, and environmental friendliness [5,24].

In this research, the dye removal capacity, the effect of contact time, pH, initial concentration of dyes, adsorbent dosage and aqueous solution temperature have been carried out. The adsorption isotherm was checked by using the Freundlich isotherm and the Langmuir isotherm. The maximum dye removal capacity was determined by the Langmuir equation. The kinetics process in the adsorption of MB and CR dyes onto  $\text{TiO}_2$ -sulfonated carbon-derived from *E. crassipes* were investigated by using Lagergren and Svenska equation. The Van't Hoff's plots were used to investigate the parameter thermodynamic enthalpy, entropy and free energy.

## 2. Materials and Methods

### 2.1 Carbonization Process

The *E. crassipes* from a pool around Mularman University were used as raw materials sources. The trunk of *E. crassipes* was washed with tap water to remove impurities. Then, it was cut and dried in oven at  $110 \text{ }^\circ\text{C}$  overnight. The carbon was derived from *E. crassipes* powder after the carbonization process in a furnace at  $600 \text{ }^\circ\text{C}$  for 1 h. The carbon-derived from *E. crassipes* was labeled as CEC.

## 2.2 Sulfonation Process

The sulfonation process follows the previous research [23]. Every gram carbon was mixed with 6 mL sulfuric acid (98%, JT Baker) and stirred at room temperature for 3 h. The acid remaining in the sample was washed with distilled water and dried in an oven at 110 °C for overnight. Furthermore, the sulfonated carbon-derived from *E. crassipes* is labeled as SCEC.

## 2.3 Titania Impregnation

The impregnation process was done by immersing 1 g sulfonated carbon in 10 mL toluene (Merck) containing titanium(IV) isopropoxide (500 μmol, Aldrich) and stirred until all of the toluene solvent completely evaporated. The residual toluene was removed from the sample by washing with ethanol (Merck) and subsequently dried at 110 °C for overnight. Then the sample was calcined at 350 °C for 2 h. The sulfonated carbon which has been impregnated is TiO<sub>2</sub>-sulfonated carbon-derived from *E. crassipes* with a denotation of TiO<sub>2</sub>/SCEC.

## 2.4 Samples Characterization

The elemental analysis, the Fourier transform infrared spectroscopy, X-ray diffraction (XRD), thermogravimetric analysis (TGA) and differential thermal analysis (DTA), scanning electron microscopy (SEM), and nitrogen adsorption-desorption studies were used to investigate the properties of TiO<sub>2</sub>/SCEC adsorbent. The elemental analysis by X-ray Fluorescence (PANalytical MiniPa 14 type) was used to determine elements contained in the adsorbent. The FTIR spectroscopy (IR-Prestige-21 Shimadzu) was used to determine the functional groups in the adsorbent. Crystallinity and phase content of the sam-

ple was investigated by XRD (Phillips PANalytical X'Pert PRO type), with the Cu Kα (λ = 1.5406 Å) radiation. The Thermal Gravimetric Analysis (TGA) data which collected using Linseis Type STA PT1600 instrument with heat rate 10 °C.min<sup>-1</sup> and target temperature 800 °C was used to determine the amount of carbon in the adsorbent. SEM images which obtained from a FEI Inspect S50 instrument was used to investigate the surface morphology of the adsorbent. The nitrogen adsorption-desorption isotherms, surface area, micropore area and pore size distribution were created from the data which collected from a Quantachrome Nova 1200e instrument.

## 2.5 Adsorption Test

The performance of adsorbent was tested using Methylene blue (C<sub>16</sub>H<sub>18</sub>ClN<sub>3</sub>S.xH<sub>2</sub>O, x = 2-3, Merck) and Congo red (C<sub>32</sub>H<sub>22</sub>N<sub>6</sub>O<sub>6</sub>S<sub>2</sub>, Merck) dyes. The Methylene blue and Congo red structures were shown in Figure 1. The adsorption was carried out by mixing 0.250 g the adsorbent with 25 mL of MB solution (100 mg.L<sup>-1</sup>) or 0.125 g adsorbent with 25 mL of CR dye solution (100 mg.L<sup>-1</sup>) in a 100 mL beaker glass at room temperature for the duration of 5, 10, 15, 20, 25, 30, 60, 120, and 240 min. Furthermore, the adsorbent was separated from the filtrate by centrifugation and the residual dye in the filtrate was analyzed by using UV-Vis spectrophotometer by monitoring the changes in absorbance at 660 nm for MB and 550 nm for CR. The effect of particle size of adsorbent for dye sorption capacity was investigated by varying size such as 40-100; 100-200 and < 200 mesh. The effect of pH was evaluated in the pH range of 2.0-10.0 for MB and 5.0-10.0 for CR dye. The initial pH of dye solution

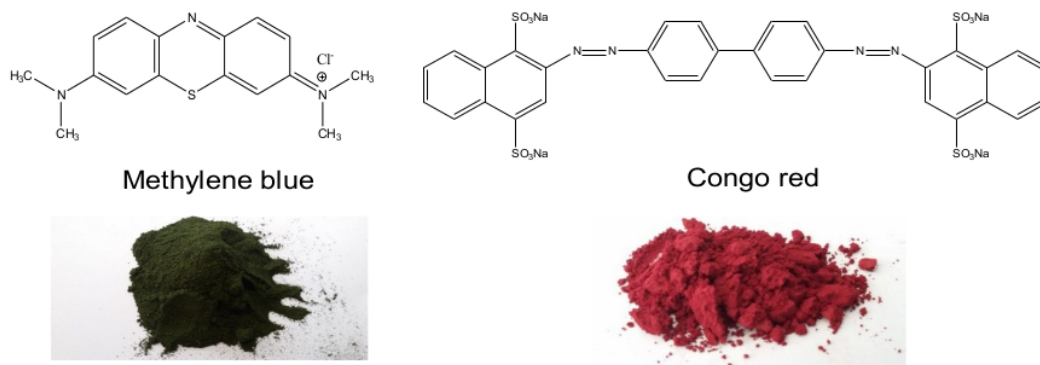


Figure 1. The chemical structure of dyes.

was fixed by the mixing 0.1 M HCl and 0.1 M NaOH. The effect of initial dye concentration was investigated by varying initial concentrations 100, 200, 300, 400, 500, and 600 mg.L<sup>-1</sup>. In this study, the thermodynamic parameters (enthalpy, entropy and free energy) also were investigated with follow the adsorption data of the effect of temperature (25, 40 and 50 °C).

The adsorption efficiency of the adsorbent was determined following Equation (1) [16,23].

$$\text{Adsorption efficiency (\%)} = \frac{(C_0 - C_t)}{C_0} \times 100\% \quad (1)$$

where  $C_0$  is the dye initial concentration and  $C_t$  is the dye concentration after adsorption time  $t$  (mg.L<sup>-1</sup>) in the solution. The adsorption capacity  $q_t$  of adsorbent (mg/g) was calculated with following the equation [25,26]:

$$q_t = \frac{(C_0 - C_t)V}{W} \quad (2)$$

and  $V$  is the volume of dye solution (mL) and  $W$  is weight adsorbent (g).

The adsorption kinetic of MB and CR adsorption onto TiO<sub>2</sub>/SCEC adsorbent were described by the pseudo-first-order kinetic models and the pseudo-second-order kinetic model. The pseudo-first-order kinetic models expression of Lagergren base on the sorption capacity of adsorbent is expressed as [27-29]:

$$\frac{dq_t}{dt} = k_1(q_e - q_t) \quad (3)$$

$$\frac{dq_t}{(q_e - q_t)} = k_1 dt \quad (4)$$

Integration of Equation (4) for the boundary condition  $t = 0$  to  $t = t$  and  $q_t = 0$  to  $q_t = t$  can be found the equation:

$$\ln(q_e - q_t) = \ln q_e - k_1 t \quad (5)$$

where  $k_1$  (g/mg.h) is the rate constant for Lagergren pseudo first-order,  $q_e$  and  $q_t$  are the amounts of dye adsorbed per gram of adsorbent (mg/g) at equilibrium and any time  $t$ . The value of  $k_1$  and  $q_e$  can be determined from slope and intercept from plot  $\ln(q_e - q_t)$  versus  $t$ .

The pseudo-second-order kinetic model is also based on the sorption capacity of the adsorbent. Linier form of pseudo-second-order kinetic model was expressed by Ho and McKay in Equations (6-7)[30,31]:

$$\frac{dq_t}{dt} = k_2(q_e - q_t)^2 \quad (6)$$

$$\frac{dq_t}{(q_e - q_t)^2} = k_2 dt \quad (7)$$

Integration of Equation (7) for the boundary condition  $t = 0$  to  $t = t$  and  $q_t = 0$  to  $q_t = t$  can be found Equation (8).

$$\frac{1}{(q_e - q_t)} - \frac{1}{q_e} = k_2 t \quad (8)$$

Equation (8) was rearranged to obtain:

$$\frac{q_t}{q_e(q_e - q_t)} = k_2 t \quad (9)$$

$$q_t = k_2 q_e^2 t - k_2 q_e q_t \quad (10)$$

Equation (10) was divided by  $k_2 q_e^2$  to obtain:

$$\frac{1}{k_2 q_e^2} = \frac{t}{q_e} - \frac{t}{q_e} \quad (11)$$

Rearrangement of equation (11) was obtained:

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e} \quad (12)$$

where  $k_2$  (g.mg<sup>-1</sup>.h<sup>-1</sup>) is the rate constant for pseudo-second-order,  $q_e$  and  $q_t$  are the amounts of dye adsorbed per gram of adsorbent (mg.g<sup>-1</sup>) at equilibrium and any time  $t$ . The value of  $k_2$  and  $q_{e,cal}$  can be calculated from intercept and slope form the plot  $t$  versus  $t/q_t$ .

The activation energy ( $E_a$ ) for adsorption process onto TiO<sub>2</sub>/SCEC adsorbent was investigated using the Arrhenius equation. The equation is given as [25,32]:

$$\ln k = \ln A - \frac{E_a}{RT} \quad (13)$$

$k$ ,  $E_a$  (kJ.mol<sup>-1</sup>),  $T$  (K),  $R$  (J.mol<sup>-1</sup>.K<sup>-1</sup>) and  $A$  are the rate constant, Arrhenius activation energy, temperature of the adsorption medium, the ideal gas constant (8.314 J.mol<sup>-1</sup>.K<sup>-1</sup>), and the Arrhenius factor, respectively. The Arrhenius activation energy was determined from slope of plotting  $\ln k$  versus  $1/T$ .

The adsorption isotherm of adsorption process onto TiO<sub>2</sub>/SCEC adsorbent was investigated by the Langmuir and the Freundlich isotherm. The Langmuir isotherm is correlated with monolayer sorption onto a surface. The Langmuir equation is given by [33-35]:

$$\frac{C_e}{q_e} = \frac{1}{Q_{max}} C_e + \frac{1}{Q_{max} b} \quad (14)$$

where  $Q_{max}$  and  $b$  are Langmuir constant which correlated with maximum adsorption capacity and bonding energy of adsorption, respectively. The slope and intercept of the linier plot of  $C_e/q_e$  versus  $C_e$  was used to determine  $Q_{max}$  and  $b$ .

The Freundlich isotherm model assumes heterogeneous site energies of sorption. The

Freundlich equation is followed as [25,34]:

$$\ln q_e = \ln K_F + \left(\frac{1}{n}\right) \ln C_e \quad (15)$$

where  $K_F$  and  $n$  are the Freundlich constant which indicate adsorption capacity and adsorption intensity, respectively. The intercept and slope of the plot of  $\ln q_e$  versus  $\ln C_e$  were used to calculate  $K_F$  ( $\text{mg}\cdot\text{g}^{-1}$ ) and  $n$  ( $(\text{L}\cdot\text{mg}^{-1})^{1/n}$ ).

The thermodynamic parameters free energy ( $\Delta G$ ), enthalpy ( $\Delta H$ ) and entropy ( $\Delta S$ ) were determined by the experiments that were created at different temperatures. The value changes of  $\Delta H$  and  $\Delta S$  of adsorption were estimated from the following equation [32,36]:

$$\Delta G^0 = -RT \ln K_C \quad (16)$$

$$\Delta G^0 = \Delta H - T\Delta S \quad (17)$$

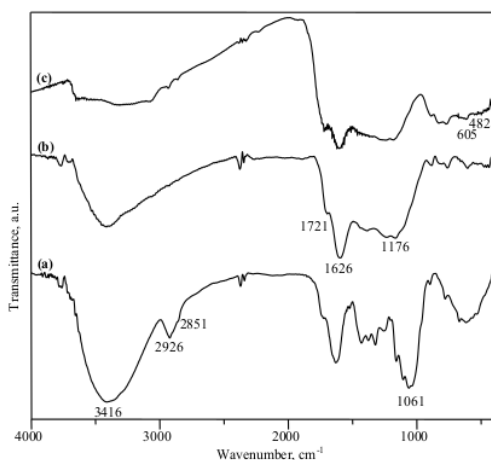
The correlation of equation (16) and (17) was obtained:

$$-RT \ln K_C = \Delta H - T\Delta S \quad (18)$$

Equation (18) was arranged to obtain Van't Hoff equation:

$$\ln K_C = -\frac{\Delta H}{R} \frac{1}{T} + \frac{\Delta S}{R} \quad (19)$$

with  $K_C$  is the equilibrium constant which calculated by the equation:



**Figure 2.** FTIR spectra of (a) CEC, (b) SCEC and (c)  $\text{TiO}_2/\text{SCEC}$  adsorbent.

**Table 1.** Elements analysis of SCEC and  $\text{TiO}_2/\text{SCEC}$  adsorbent by using XRF.

Sample	Element concentration (wt%)														
	Si	P	S	K	Ca	Ti	Mn	Fe	Ni	Cu	Zn	Br	Mo	Eu	Yb
SCEC	2.90	1.00	9.00	1.2	48.70	0.33	4.49	2.40	0.20	0.48	0.27	1.20	25.00	0.40	0.40
$\text{TiO}_2/\text{SCEC}$	1.10	0.55	3.40	0.35	11.10	73.30	1.50	0.81	0.05	0.22	0.19	0.60	6.10	0.20	0.30

$$K_C = \frac{C_1}{C_2} \quad (20)$$

where  $T$ ,  $R$ ,  $C_1$  and  $C_2$  are temperature (K), the gas constant ( $8.314 \text{ J}\cdot\text{K}^{-1}\cdot\text{mol}^{-1}$ ), the quantity of methylene blue dye adsorbed per unit mass of adsorbent and the concentration of methylene blue dye in aqueous phase, respectively. The linear plot of  $\ln K_C$  versus  $1000/T$  was used to determine of  $\Delta S$  and  $\Delta H$  which calculated from the slope and intercept. The positive value of  $\Delta S$  indicates that the increase in randomness of ongoing process. The negative value of  $\Delta H$  indicates that the adsorption process is exothermic in nature and otherwise, the positive value of  $\Delta H$  indicates that the adsorption process is endothermic in nature. Furthermore, the value of  $\Delta G$  was calculated with follow the Equation (17). The feasibility and spontaneity of adsorption process was shown by the negative value of  $\Delta G$  at each temperature.

### 3. Results and Discussion

#### 3.1 Physical Properties

The XRF results show that the SCEC and  $\text{TiO}_2/\text{SCEC}$  adsorbents consist of major elements, such as: calcium (Ca), silicon (Si), manganese (Mn), and molybdenum (Mo). Many transition metals elements also were found inside the both adsorbents like ferum (Fe), copper (Cu), nickel (Ni), and zinc (Zn). The effect of sulfonation process onto CEC was indicated by percentage of sulfur (9.0%) and it decreases after titanium impregnation process. The percentage of titanium into  $\text{TiO}_2/\text{SCEC}$  adsorbents reach 73.30% after impregnation process by titanium(IV) isopropoxide (500  $\mu\text{mol}$ , Aldrich). The complete list of elements is displayed in Table 1.

Figure 2 shows the FTIR spectra for (a) CEC, (b) SCEC, and (c)  $\text{TiO}_2/\text{SCEC}$ . The O-H stretching mode of the -COOH and phenolic OH groups on the samples were indicated by the broad band at  $3416 \text{ cm}^{-1}$  in the FTIR spectra [37]. The C-H (aliphatic) stretching vibration in the sample was investigated by the adsorption peaks at 2926 and 2851  $\text{cm}^{-1}$ . The vibration band at  $1626 \text{ cm}^{-1}$  indicates as the adsorption peak of aromatic C=C stretching mode

in polyaromatics. The respective intensity of O–H, C–H, and C=C vibration decrease after sulfonation and TiO<sub>2</sub> impregnation onto CEC. The effect of sulfonation process on the adsorbent was proven by the attachment of –SO<sub>3</sub>H groups which indicated by the adsorption band at 1177 and 605 cm<sup>-1</sup> [38]. The effect of titanium impregnation on the adsorbent can be proven by the peak at around 482 cm<sup>-1</sup> is attributed symmetric O–Ti–O stretching that caused the vibration of Ti–O bond.

The low crystallinity of CEC, SCEC, and TiO<sub>2</sub>/SCEC adsorbent were shown by the XRD pattern in Figure 3. The broad peak at 2θ value of 10–30° (C (002)) attributed to amorphous car-

bon composed of aromatic carbon sheets, can be observed in all adsorbents [37,39,40]. Base on JCPDS number 41-1487, a diffraction peak at 2θ value of 35–50° C (101) indicated that all adsorbents formed turbostratic graphite structure [37,41]. Base on JCPDS number 00-004-0477, the crystal structure of TiO<sub>2</sub> can be observed from the dominant peaks at 25.4° and other peaks at 2θ = 37.9, 48.1, 54.1 and 55.2° but does not appear in Figure 3 (c) due to the amount of titania that added onto SCEC very small (500 μmol).

The investigation of weight loss in the TiO<sub>2</sub>/SCEC adsorbent which used to control the amount of carbon while increasing the temperature can be shown from the TGA and DTA curves in Figure 4. The weight loss of the adsorbent was obtained in one step. The amount of carbon (18 mg) lose when the temperature increase start at 0 °C, reached midpoint at 400 °C and continued until 800 °C.

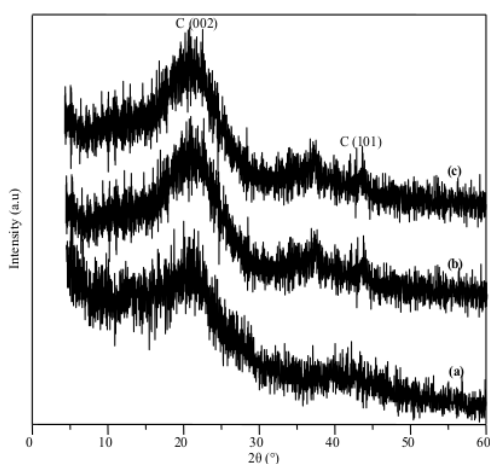


Figure 3. XRD pattern of (a) CEC, (b) SCEC and (c) TiO<sub>2</sub>/SCEC adsorbent.

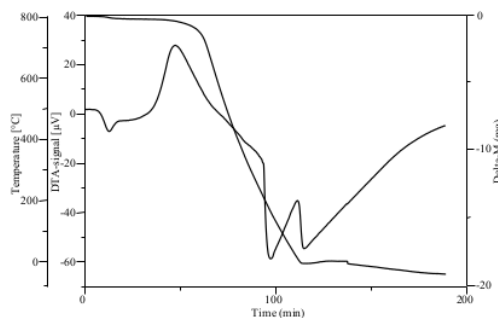


Figure 4. TGA and DTA curves of TiO<sub>2</sub>/SCEC adsorbent.

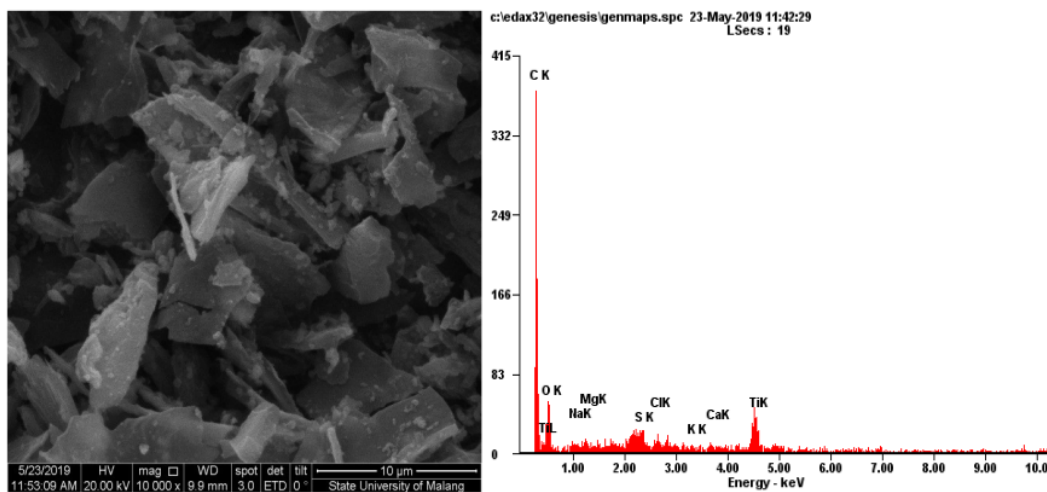
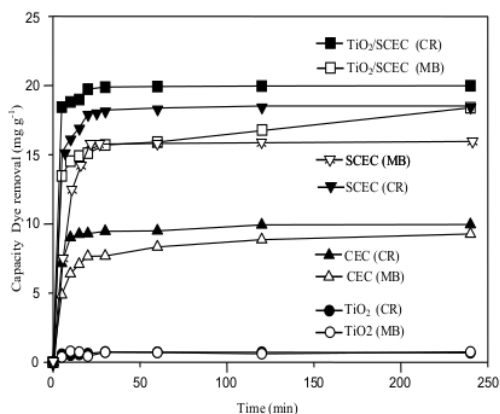


Figure 5. SEM/EDX Image of TiO<sub>2</sub>/SCEC adsorbent.

The SEM image of  $\text{TiO}_2/\text{SCEC}$  adsorbent was shown in Figure 5. The SEM image shows that the  $\text{TiO}_2/\text{SCEC}$  adsorbent is in the fiber form, which is arranged as fibrils. Hollow spaces started to appear after the *E. crassipes* was carbonated and sulfonated. The surface morphology of  $\text{TiO}_2/\text{SCEC}$  adsorbent is roughness. The EDX results (Table 2) show that many elements C, O, Na, Mg, S, Cl, K, Ca, and Ti were detected inside  $\text{TiO}_2/\text{SCEC}$  adsorbent. The adsorbent consists of carbon as a major element. The effect of the sulfonation and titania impregnation process onto adsorbent was proven by the S and Ti elements were found inside adsorbent.

The other physical properties the BET surface area, pore volume and mean pore size of CEC, SCEC, and  $\text{TiO}_2/\text{SCEC}$  were obtained from the nitrogen adsorption-desorption analysis. The complete data were shown in Table 3. The BET surface area, pore volume and mean pore size showed the following values; 94.8  $\text{m}^2\cdot\text{g}^{-1}$ , 0.109  $\text{cm}^3\cdot\text{g}^{-1}$ , 2.303 nm for CEC, 97.8  $\text{m}^2\cdot\text{g}^{-1}$ , 0.184  $\text{cm}^3\cdot\text{g}^{-1}$ , 3.770 nm for SCEC and 233.0  $\text{m}^2\cdot\text{g}^{-1}$ , 0.248  $\text{cm}^3\cdot\text{g}^{-1}$ , 2.131 nm for  $\text{TiO}_2/\text{SCEC}$ , respectively. It can be seen that the BET surface area and pore volume increase can be increased by sulfonation process. And BET surface area and pore volume increase drastically when SCEC was modified to be  $\text{TiO}_2/\text{SCEC}$  by impregnation process.



**Figure 6.** Effect of contact time on the amount of MB and CR dyes removal on  $\text{TiO}_2$ ; CEC, SCEC and  $\text{TiO}_2/\text{SCEC}$ . Conditions: (pH: 6.9, concentration MB and CR 100  $\text{mg}\cdot\text{L}^{-1}$ , weight adsorbent for MB 250 mg and CR 125 mg, at 25°C).

### 3.2. Effect of Contact Time

The effect of contact time on the dye removal capacity ( $q_t$ ) of MB and CR onto  $\text{TiO}_2$ , CEC, SCEC and  $\text{TiO}_2/\text{SCEC}$  adsorbents are shown in Figure 6.  $\text{TiO}_2$ , CEC, and SCEC adsorbents were used as a comparison to  $\text{TiO}_2/\text{SCEC}$ . These results showed the dye removal capacity of all adsorbents to CR higher than MB. The dye removal capacity for MB or CR dye using  $\text{TiO}_2$  adsorbent almost the same  $\sim 0.7 \text{ mg}\cdot\text{g}^{-1}$  in adsorption duration time 5 to 240 min. The dye removal capacity of CEC adsorbent for CR reached 7.1  $\text{mg}\cdot\text{g}^{-1}$  in adsorption duration time 5 min and drastically increased to 9.0  $\text{mg}\cdot\text{g}^{-1}$  when adsorption duration time 10 min. The equilibrium CR adsorption was reached within 30 min and the CR removal capacity increased with increasing contact time. The MB removal capacity of CEC adsorbent reached 4.8  $\text{mg}\cdot\text{g}^{-1}$  within adsorption duration time 5 min and increased up to 6.3  $\text{mg}\cdot\text{g}^{-1}$  after 10 min. Furthermore, after adsorption duration time 120 min the equilibrium MB achieved 8.9  $\text{mg}\cdot\text{g}^{-1}$ . The CR removal capacity of SCEC achieved 15.1  $\text{mg}\cdot\text{g}^{-1}$  within adsorption duration time 5 min and increased up to 16.1  $\text{mg}\cdot\text{g}^{-1}$  after 10 min. Then, after adsorption duration time 30 min the equilibrium CR reached  $\sim 18 \text{ mg}\cdot\text{g}^{-1}$ . The MB removal capacity of SCEC achieved  $\sim 9.6 \text{ mg}\cdot\text{g}^{-1}$  within adsorption duration time 5 min

**Table 2.** Physical properties of SEM EDX of the adsorbent.

Element	wt%
C	68.9
O	20.3
Na	0.6
Mg	0.7
S	1.4
Cl	0.8
K	0.2
Ca	0.7
Ti	6.3

**Table 3.** Physical properties of CEC, SCEC and  $\text{TiO}_2/\text{SCEC}$  adsorbents

Samples	BET surface area ( $\text{m}^2/\text{g}$ )	Pore Volume ( $\text{cm}^3/\text{g}$ )	Mean pore size (mm)
CEC	94.8	0.109	2.302
SCEC	97.78	0.184	3.770
$\text{TiO}_2/\text{SCEC}$	233.0	0.248	2.131



and after 30 min the equilibrium MB  $\sim 16$   $\text{mg}\cdot\text{g}^{-1}$  was achieved. The CR removal capacity of  $\text{TiO}_2/\text{SCEC}$  achieved  $18.4$   $\text{mg}\cdot\text{g}^{-1}$  within adsorption duration time 5 min and it showed a slight increase of the equilibrium,  $19.9$   $\text{mg}\cdot\text{g}^{-1}$  after 30 min. The MB removal capacity achieved  $7.8$   $\text{mg}\cdot\text{g}^{-1}$  within adsorption duration time 5 min and increased to a certain extent showing that the equilibrium reached  $9.9$   $\text{mg}\cdot\text{g}^{-1}$  after 30 min. Base on the effect of contact time show that CEC, SCEC and  $\text{TiO}_2/\text{SCEC}$  adsorbents more easy to adsorb CR than MB dyes. The sulfonation process onto CEC and impregnation  $\text{TiO}_2$  onto SCEC support give positive effect to increase the dye re-

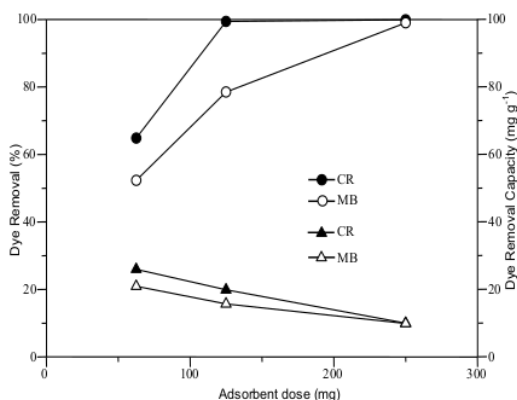
moval capacity which was proven by the dye removal capacity of  $\text{TiO}_2/\text{SCEC} > \text{SCEC} > \text{CEC} > \text{TiO}_2$  for both dyes. The increasing dye removal capacities were caused by the increasing surface area.

### 3.3. Effect of Dose

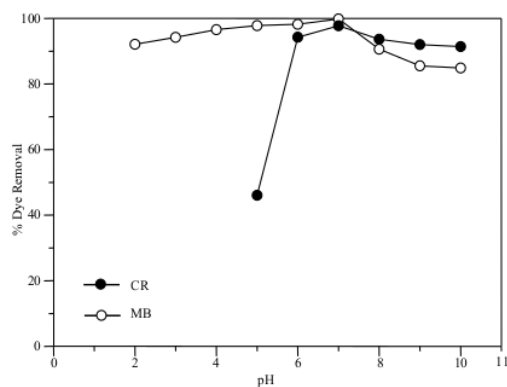
The plot between the adsorbent dose (mg) versus the percentage of dye removal and dye removal capacity was shown in Figure 7. In this experiment, the dye removal capacity decreased from  $20.93$  to  $9.9$   $\text{mg}\cdot\text{g}^{-1}$  for MB and  $25.94$  to  $9.99$   $\text{mg}\cdot\text{g}^{-1}$  for CR when adsorbent dose has increased from  $62.5$  to  $250$  mg. An increase in the dose of adsorbent is not necessary followed by an increase of the dye removal capacity owing to the ratio adsorbent dose to dye concentration is high causing the amount of active sites remaining not covered by dye molecules also high. This result almost in line with the previous research that investigated methylene blue adsorption onto activated carbon prepared from cashew nut shell [3]. The adsorbent dose optimum was achieved  $250$  mg for MB and  $125$  mg for CR when the percentage dyes removal almost  $\sim 100\%$  for both dyes.

### 3.4 Effect of pH

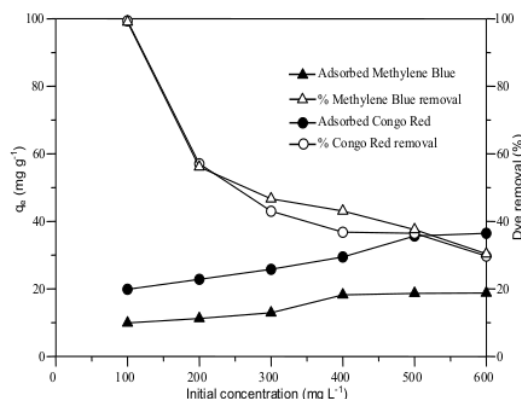
The effect of pH on the MB and CR dyes adsorption onto  $\text{TiO}_2/\text{SCEC}$  adsorbent was shown in Figure 8. The removal of MB dye from aqueous solution was effective in the pH range of 2–7. The removal of MB at pH less than 2 was ineffective as the effect of the excess  $\text{H}^+$  ion that may caused competition between the positively



**Figure 7.** Effect of  $\text{TiO}_2/\text{SCEC}$  adsorbent dose on sorption of MB and CR. Conditions: (pH: 6.9, concentration MB and CR from  $100$   $\text{mg}\cdot\text{L}^{-1}$ , time 30 min, weight adsorbent:  $62.5$ ,  $125$  and  $250$  mg, at  $25^\circ\text{C}$ ).



**Figure 8.** Effect of pH on % MB and CR dye removal on  $\text{TiO}_2/\text{SCEC}$ . Conditions: (pH: 6.9, concentration MB and CR from  $100$   $\text{mg}\cdot\text{L}^{-1}$ , weight adsorbent for MB  $250$  mg and CR  $125$  mg, at  $25^\circ\text{C}$ ).



**Figure 9.** Effect of initial dye concentration on sorption of MB and CR on  $\text{TiO}_2/\text{SCEC}$ . Conditions: (pH: 6.9, concentration MB and CR from  $100$ – $600$   $\text{mg}\cdot\text{L}^{-1}$ , weight adsorbent for MB  $250$  mg and CR  $125$  mg, at  $25^\circ\text{C}$ ).

charged sites ( $H^+$ ) and the dye<sup>+</sup> in adsorption process. The removal of MB tends to decrease due to the excess of the number of negatively charged sites ( $OH^-$ ) at pH value more than 8 [3]. Otherwise, the CR removal was beneficial in the pH range of 6–10. The removal of CR was not effective at pH less than 5.

### 3.5 Effect of Initial Concentration

The effect of different initial MB and CR dyes concentrations on the adsorption process is shown in Figure 9. The dyes removal capacity of both dyes increased when the initial concentration increased. At equilibrium, the MB removal capacity increased from 9.9 to 18.8  $mg.g^{-1}$  and the CR removal capacity 19.9 to 36.4  $mg.g^{-1}$  when the initial concentration of both dyes increased from 100 to 600  $mg.L^{-1}$ . The increasing of the amount of dye removal at equilibrium condition was caused by an increasing the initial concentration that can increase the driving force to reduce all mass transfer resistances of the MB and CR dyes molecules between the aqueous and solid phase of the adsorbent [2]. Otherwise, the percentage of both dyes adsorbed decrease from 99.0 to 30.4% for MB and 99.4 to 29.7% for CR when the initial concentration increased from 100 to 600  $mg.L^{-1}$ .

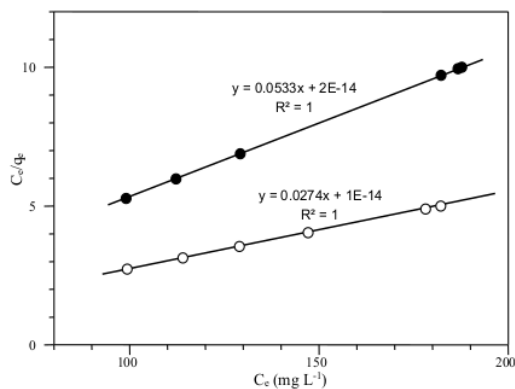


Figure 10. The Langmuir isotherm of MB and CR onto  $TiO_2/SCEC$ .

### 3.6 Adsorption Isotherms

The Langmuir and the Freundlich adsorption isotherms were used to describe the interaction between the dyes and the  $TiO_2/SCEC$  adsorbent shown in Figure 10 and 11, respectively. For the removal of MB and CR red by  $TiO_2/SCEC$  adsorbent, 25 mL of various dyes concentrations (between 100 and 600  $mg.L^{-1}$ ) were mixed with 0.250 mg sorbent for MB and 0.125 g sorbent for CR, then stirred at 300 rpm for 30 min at room temperature. The results are shown in Table 4.

From the Langmuir isotherm can be obtained the value of the constants of  $Q_{max}$ ,  $b$  and  $R^2$  as 36.5  $mg.g^{-1}$ ,  $2.74 \times 10^{12}$  and 1.000 for CR and 18.8  $mg.g^{-1}$ ,  $2.66 \times 10^{12}$  and 1.000 for MB, respectively. Base on the value of maximum adsorption capacity ( $Q_{max}$ ), bonding energy ( $b$ ) can be used to conclude that adsorption of CR was easier than MB onto  $TiO_2/SCEC$  adsorbent.

The constant  $K_F$ ,  $n$ , and  $R^2$  were determined from the Freundlich isotherm was obtained 0.2017, 1.002, and 1.000 for CR and 0.1002, 1.001, and 1.000 for MB, respectively. Base on  $K_F$  value shows that the MB and CR adsorption process was the heterogeneous adsorption process on the surface of on  $TiO_2/SCEC$  adsor-

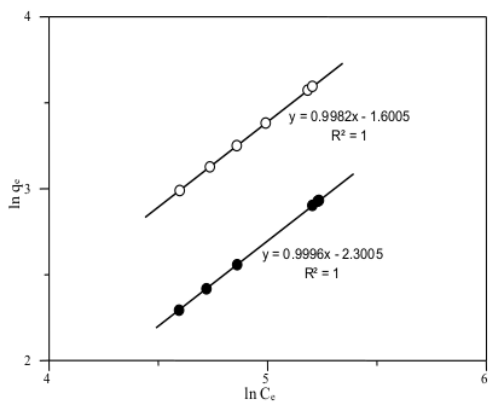


Figure 11. The Freundlich isotherm of MB and CR onto  $TiO_2/SCEC$ .

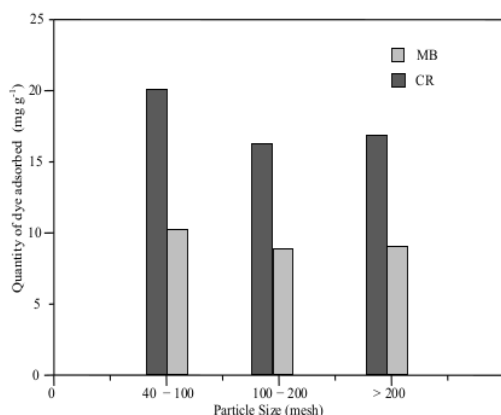
Table 4. Langmuir and Freundlich Isotherm models for CR and MB dyes adsorption on  $TiO_2/SCEC$  adsorbent.

Dye removal	Langmuir Isotherm			Freundlich Isotherm		
	$Q_0$	$b$	$R^2$	$K_F$	$n$	$R^2$
Congo Red (CR)	36.5	$2.74 \times 10^{12}$	1.000	0.2017	1.002	1.000
Methylene Blue (MB)	18.8	$2.66 \times 10^{12}$	1.000	0.1002	1.001	1.000

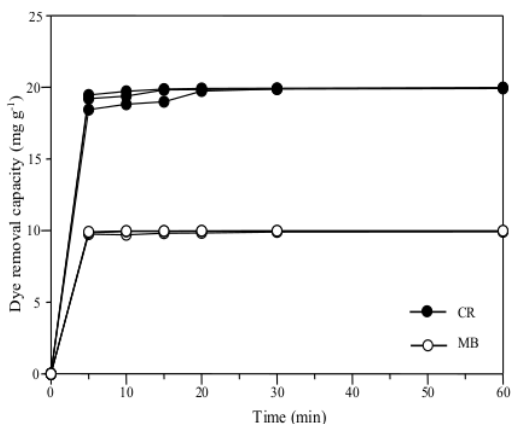
bent which was done through a multi-layer adsorption mechanism. The  $n$  values of MB and CR adsorption process are more than 1, which indicates favorable adsorption of both dyes onto the  $\text{TiO}_2/\text{SCEC}$  adsorbent.

### 3.7 Effect of Particle Size

The effect of adsorbent particle size on dyes removal capacity is shown in Figure 12. The effect of  $\text{TiO}_2/\text{SCEC}$  particle size batch contact experiments were carried out with three parti-



**Figure 12.** Effect of particle size of adsorbent on the removal of MB and CR dyes ( $\text{mg.g}^{-1}$ ) on  $\text{TiO}_2/\text{SCEC}$ . Conditions: (pH: 6.9, concentration MB and CR from  $100 \text{ mg.L}^{-1}$ , weight adsorbent for MB 250 mg and CR 125 mg, at  $25^\circ\text{C}$ ).

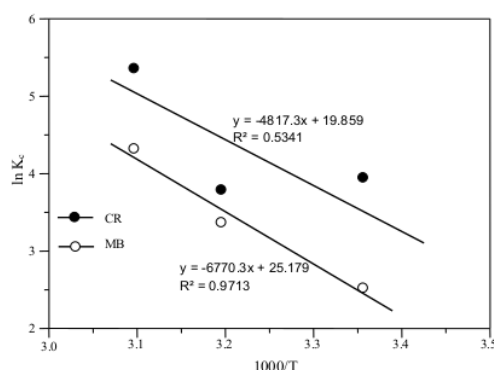


**Figure 13.** Effect of temperature on on the removal of MB and CR dyes ( $\text{mg.g}^{-1}$ ) on  $\text{TiO}_2/\text{SCEC}$ . Conditions: (pH: 6.9, concentration MB and CR from  $100 \text{ mg.L}^{-1}$ , weight adsorbent for MB 250 mg and CR 125 mg, at 25, 40, and  $50^\circ\text{C}$ ).

cles sizes using adsorbent mass 250 mg for MB and 125 mg for CR, the same initial concentration  $100 \text{ mg.L}^{-1}$ , and at room temperature. Three selected particle size that used in this experiment 40–100, 100–200 and  $>200$  mesh. The dyes removal capacity were 8.5, 8.7, and  $9.9 \text{ mg.g}^{-1}$  for MB and 15.9, 16.4, and  $19.7 \text{ mg.g}^{-1}$  for CR, using the above mentioned particle sizes, respectively. It is worth noting that the dyes removal capacity decreased with an increase in particle size of  $\text{TiO}_2/\text{SCEC}$  adsorbent. This phenomenon can be explained by the smaller particles size for the same amount of adsorbent has larger total surface area [27]. This experiment result was similar with the previous research done on adsorption of Congo red from aqueous solution using roots of *E. crassipes* [16].

### 3.8 Effect of Temperature

The effect of temperature on the dye removal capacity is shown in Figure 13. The dyes removal capacity at varying temperatures increased from  $9.75$  to  $9.92 \text{ mg.g}^{-1}$  (at room temperature),  $9.84$  to  $9.96 \text{ mg.g}^{-1}$  (at  $40^\circ\text{C}$ ) and  $9.91$  to  $9.98 \text{ mg.g}^{-1}$  (at  $50^\circ\text{C}$ ) for MB and  $18.44$  to  $19.92 \text{ mg.g}^{-1}$  (at room temperature),  $19.19$  to  $19.91 \text{ mg.g}^{-1}$  (at  $40^\circ\text{C}$ ) and  $19.46$  to  $19.98 \text{ mg.g}^{-1}$  (at  $50^\circ\text{C}$ ) for CR. It can be clearly seen that by increasing the temperature, the dye removal capacity of both dyes also increased slowly and reached optimum at almost  $9.98 \text{ mg.g}^{-1}$  for MB and  $19.98 \text{ mg.g}^{-1}$  for CR, respectively. The increasing of dye removal capacity when the adsorption temperature increase was as evident that the adsorption of both dyes onto  $\text{TiO}_2/\text{SCEC}$  adsorbent was endothermic in nature.



**Figure 14.** Van't Hoff's plots,  $\ln Kc$  vs.  $1000/T$  for CR and MB dyes adsorption on  $\text{TiO}_2/\text{SCEC}$  adsorbent

Figure 14 shows Van't Hoff's plots that used to determine the thermodynamic parameter value of free energy ( $\Delta G$ ), enthalpy ( $\Delta H$ ), and entropy ( $\Delta S$ ). The  $\Delta H$  and  $\Delta S$  are calculated based on the slope and intercept. Furthermore,  $\Delta H$  and  $\Delta S$  were used to determine  $\Delta G$  according to equation (17). The values of  $\Delta H$ ,  $\Delta S$  were  $+56.56 \text{ kJ.mol}^{-1}$ ,  $+0.229 \text{ kJ.mol}^{-1}.\text{K}^{-1}$  for MB and  $+40.05 \text{ kJ.mol}^{-1}$ ,  $+0.178 \text{ kJ.mol}^{-1}.\text{K}^{-1}$  for CR, respectively. The average values of  $\Delta G$  were  $-14.84 \text{ kJ.mol}^{-1}$  for MB and  $-15.52 \text{ kJ.mol}^{-1}$  for CR. Positive value of  $\Delta H$  indicates that the adsorption process of both dyes onto  $\text{TiO}_2/\text{SCEC}$  adsorbent was endothermic in nature. The increasing in degree of system irregularities was identified by the positive value of  $\Delta S$ . The feasibility and spontaneity of ongoing dyes adsorption process onto  $\text{TiO}_2/\text{SCEC}$  adsorbent was shown by the negative value of  $\Delta G$ . The complete data of the thermodynamic parameters are listed in Table 5.

### 3.9 Adsorption of Kinetic

The kinetics data obtained from MB and CR dyes adsorption experiment onto  $\text{TiO}_2/\text{SCEC}$  adsorbent were investigated using pseudo-first-order and pseudo-second order kinetics models. The complete results are given in Table 6. Based on the experiment results, the dyes removal onto  $\text{TiO}_2/\text{SCEC}$  adsorbent at temperature from 25 to 50 °C did not obey a first-order reaction model due to the correlation coefficients  $R^2 \ll 1.000$  and the dyes removal capacity calculated

( $q_{e,cal}$ )  $\ll$  the dyes removal capacity experimental ( $q_{e,exp} = 9.976 \text{ mg.g}^{-1}$  for MB;  $19.956 \text{ mg.g}^{-1}$  for CR).

Otherwise, the parameters kinetic models in Table 6 showed good compliance with the pseudo-second order. This was proven by the correlation coefficients ( $R^2$ ) for linear plots for MB and CR removal on to  $\text{TiO}_2/\text{SCEC}$  adsorbent were 1.000. The rate constant ( $k_2$ ) value of the removal of the dyes onto  $\text{TiO}_2/\text{SCEC}$  adsorbent increased with increasing temperature for both dyes. This phenomenon is in line with previous researches that have been reported in the reference for different adsorbate dyes [42]. The highest rate constant of pseudo-second order was found to be 3.7110 for MB and 0.3608 for CR at 50 °C. The calculated ( $q_{e,cal}$ ) and experimental ( $q_{e,exp}$ ) dyes removal capacity at 50 °C were found to be 9.990 and 9.986  $\text{mg.g}^{-1}$  for MB and 20.004 and 19.981  $\text{mg.g}^{-1}$  for CR. This data indicated that the dyes removal capacity calculated ( $q_{e,cal}$ ) was close to the dyes removal capacity experimental ( $q_{e,exp}$ ). The same phenomenon was found in experiments at 25 °C and 40 °C. The experimental data for dyes removal on to on to  $\text{TiO}_2/\text{SCEC}$  adsorbent fit the pseudo-second order kinetic model. Similar kinetic applications were also determined for dyes removal from aqueous solution with activated carbon produced from tea seed shells [36], activated carbon from cashew nut shell [3], in which the adsorption rate of dyes removal were described by a pseudo-second order.

**Table 5.** Thermodynamic parameters data for CR and MB dyes adsorption on  $\text{TiO}_2/\text{SCEC}$  adsorbent

Dyes	$\Delta H$ (kJ/mol)	$\Delta S$ (kJ/mol K)	$\Delta G$ (kJ/mol)			$E_a$ (kJ/mol)
			25 °C	40 °C	50 °C	
MB	+56.564	+0.229	-11.779	-15.219	-17.513	82.9
CR	+40.051	+0.178	-13.137	-15.814	-17.599	33.4

**Table 6.** First order and Pseudo second order kinetics and Freundlich isotherm constants for CR and MB dyes adsorption on  $\text{TiO}_2/\text{SCEC}$  adsorbent.

Dye removal	Temperature (°C)	First Order			Pseudo Second Order			$q_{e,exp}$ (mg.g <sup>-1</sup> )
		$q_{e,cal}$ (mg.g <sup>-1</sup> )	$k_1$ (g.mg <sup>-1</sup> .h <sup>-1</sup> )	$R^2$	$q_{e,cal}$ (mg.g <sup>-1</sup> )	$k_2$ (g.mg <sup>-1</sup> .h <sup>-1</sup> )	$R^2$	
Congo Red (CR)	25	0.932	0.0371	0.7851	20.000	0.1276	1.000	19.976
	40	1.153	0.1226	0.8717	20.000	0.2475	1.000	19.910
	50	0.636	0.0953	0.8887	20.004	0.3608	1.000	19.981
Methylene Blue (MB)	25	0.540	0.0295	0.5640	9.980	0.2997	1.000	9.976
	40	0.123	0.1451	0.9729	9.920	2.3090	1.000	9.965
	50	0.143	0.1728	0.9456	9.990	3.7110	1.000	9.986

#### 4. Conclusions

Many research using *E. crassipes* as the raw material of adsorbents to remove dyes have been conducted. In this work, the *E. crassipes* is a tropical plant, which is low in cost and abundant elsewhere. Therefore, it was chosen as an alternative, natural source of carbon and followed by sulfonation and titania impregnation to be an adsorbent. The batch system was used for the adsorption process of MB and CR onto TiO<sub>2</sub>/SCEC at initial concentration of 100 mg/L at room temperature. The TiO<sub>2</sub>/SCEC adsorbent exhibited higher removal capacity toward MB and CR dyes. The experiment data indicated that the adsorption kinetics of dyes onto the TiO<sub>2</sub>/SCEC followed pseudo-second order model. The parameters thermodynamic data indicated the dyes adsorption process onto the TiO<sub>2</sub>/SCEC was endothermic and spontaneous in nature. Results showed the removal of CR more effective than MB onto the TiO<sub>2</sub>/SCEC adsorbent.

#### Acknowledgements

The authors gratefully acknowledge research grant from IsDB TA. 2019, (No: 137/UN17.11/PL/2019) and education faculty and teaching training Mulawarman University, East Kalimantan Province, Indonesia.

#### References

- [1] Mahmoodi, N.M., Salehi, R., Arami, M. (2011). Binary system dye removal from colored textile wastewater using activated carbon: Kinetic and isotherm studies. *Desalination*, 272, 187-195. DOI: 10.1016/j.desal.2011.01.023.
- [2] Wanyonyi, W.C., Onyari, J.M., Shiundu, P.M. (2013). Adsorption of Methylene Blue Dye from Aqueous Solution Using Eichhornia crassipes. *Bull. Environ. Contam. Toxicol.*, 91(3), 362-366. DOI: 10.1007/s00128-013-1053-0.
- [3] Kumar, P.S., Ramalingam, S., Sathishkumar, K. (2011). Removal of methylene blue dye from aqueous solution by activated carbon prepared from cashew nut shell as a new low-cost adsorbent. *Korean J. Chem. Eng.*, 28(1), 149-155. DOI: 10.1007/s11814-010-0342-0.
- [4] Nigam, P., Armour, G., Banat, I.M., Singh, D., Marchant, R. (2000). Physical removal of textile dyes from effluents and solid-state fermentation of dye-adsorbed agricultural residues. *Bioresour. Technol.*, 72, 219-226. DOI: 10.1016/S0960-8524(99)00123-6.
- [5] Wawrzekiewicz, M., Polska-Adach, E., Hurbicki, Z. (2019). Application of titania based adsorbent for removal of acid, reactive and direct dyes from textile effluents. *Adsorption*, 25, 621-630. DOI: 10.1007/s10450-019-00062-0.
- [6] Hachem, C., Bocquillon, F., Zahraa, O., Bouchy, M. (2001). Decolorization of textile industry wastewater by the photocatalytic degradation process. *Dyes and Pigments*, 49, 117-125. DOI: 10.1016/S0143-7208(01)00014-6.
- [7] Park, C., Lee, M., Lee, B., Kim, S.-W., Chase, H.A., Lee, J., Kim, S. (2007). Biodegradation and biosorption for decolorization of synthetic dyes by *Funalia trogii*. *Biochem. Eng. J.*, 36, 59-65. DOI: 10.1016/j.bej.2006.06.007.
- [8] Crini, G. (2006). Non-conventional low-cost adsorbents for dye removal: A review. *Bioresour. Technol.*, 97, 1061-1085. DOI: 10.1016/j.biortech.2005.05.001.
- [9] Alinsafi, A., Khemis, M., Pons, M. N., Leclerc, J. P., Yaacoubi, A., Benhammou, A., Nejmeddine, A. (2005). Electro-coagulation of reactive textile dyes and textile wastewater. *Chem. Eng. Process.*, 44, 461-470. DOI: 10.1016/j.ccep.2004.06.010.
- [10] Karaoglu, M.H., Dogan, M., Alkan, M. (2010). Removal of Reactive Blue 221 by Kaolinite from Aqueous Solution. *Ind. Eng. Chem. Res.*, 49, 1534-1540. DOI: 10.1021/ie9017258.
- [11] Peng, Q., Yu, F., Huang, B., Huang, Y. (2017). Carbon-containing bone hydroxyapatite obtained from tuna fish bone with high adsorption performance for Congo red. *The Royal Soc. Chem.*, 17, 26968. DOI: 10.1039/C6RA27055G
- [12] Wang, L. (2012). Application of activated carbon derived from waste bamboo culm for the adsorption of azo disperse dye: kinetic, equilibrium and thermodynamic studies. *J. Environ. Manag.*, 102, 79-87. DOI: 10.1016/j.jenvman.2012.02.019.
- [13] Kusumawardani, R., Nurhadi, M., Wirhanuddin, Gunawan, R., Nur, H. (2019). Carbon-containing Hydroxyapatite Obtained from Fish Bone as Low-cost Mesoporous Material for Methylene Blue Adsorption. *Bull. Chem. Reac. Eng. Catal.*, 14(3), 660-671. DOI: 10.9767/brec.14.3.5365.660-671.
- [14] Muthukumar, M., Sevakumar, N. (2004). Studies on the effect of inorganic salts on decoloration of acid dye effluent by ozonation. *Dyes and Pigments*, 62, 221-228. DOI: 10.1016/j.dyepig.2003.11.002.
- [15] Sikaily, A.E., Khaled, A., Nemr, A.E., Abdelwahab, O. (2006). Removal of Methylene Blue from aqueous solution by marine green alga

- Ulva lactuca. *Chem. Ecol.*, 22, 149-157. DOI: 10.1080/02757540600579607.
- [16] Wanyonyi, W.C., Onyari, J.M., Shiundu, P.M. (2014). Adsorption of Congo Red Dye from Aqueous Solution Using Roots of Eichhornia Crassipes: Kinetic and Equilibrium Studies. *Energy Procedia*, 50, 862-869. DOI: 10.1016/j.egypro.2014.06.105.
- [17] Zawahry, M.M.E., Kamel, M.M. (2004). Removal of azo and anthraquinone dyes from aqueous solutions by Eichhornia Crassipes. *Water Research*, 38, 2967-2972. DOI: 10.1016/S0043-1354(01)00526-7.
- [18] Belessi, V., Romanos, G., Boukos, N., Lambropoulou, D., Trapalis, C. (2009). Removal of Reactive Red 195 from aqueous solutions by adsorption on the surface of TiO<sub>2</sub> nanoparticles. *J. Hazard. Mater.*, 170, 836-844. DOI: 10.1016/j.jhazmat.2009.05.045.
- [19] Wang, R., Cai, X., Shen, F. (2014). TiO<sub>2</sub> hollow microspheres with mesoporous surface: Superior adsorption performance for dye removal. *Applied Surface Sci.*, 305, 352-358. DOI: 10.1016/j.apsusc.2014.03.089.
- [20] Sriprang, P., Wongnawa, S., Sirichote, O. (2014). Amorphous titanium dioxide as an adsorbent for dye polluted water and its recyclability. *J. Sol-Gel Sci. Technol.*, 71, 86-95. DOI: 10.1007/s10971-014-3327-3.
- [21] Kamal, T., Anwar, Y., Khan, S. B., Chani, M. T. S., Asiri, A. M. (2016). Dye adsorption and bactericidal properties of TiO<sub>2</sub>/chitosan coating layer. *Carbohydrate Polym.*, 148, 153-160. DOI: 10.1016/j.carbpol.2016.04.042.
- [22] Li, J., Feng, J., Yan, W. (2013). Synthesis of Polypyrrole-Modified TiO<sub>2</sub> Composite Adsorbent and Its Adsorption Performance on Acid Red G. *J. Appl. Polym. Sci.*, 128, 3231-3239. DOI: 10.1002/app.38525.
- [23] Nurhadi, M., Widiyowati, I. I., Wirhanuddin, Chandren, S. (2019). Kinetic of Adsorption Process of Sulfonated Carbon-derived from Eichhornia crassipes in the Adsorption of Methylene Blue Dye from Aqueous Solution. *Bull. Chem. React. Eng. Catal.*, 14(1), 17-27. DOI: 10.9767/bcrec.14.1.2548.17-27.
- [24] Ito, S., Kon, Y., Nakashima, T., Hong, D., Konno, H., Ino, D., Sato, K. (2019). Titania-Catalyzed H<sub>2</sub>O<sub>2</sub> Thermal Oxidation of Styrenes to Aldehydes. *Molecules*, 24, 1-9. DOI: 10.3390/molecules24142520.
- [25] Pathania, D., Arush, S., Siddiqi, Z.M., (2016). Removal of congo red dye from aqueous system using Phoenix dactylifera seeds. *J. Mol. Liq.*, 219, 359-367. DOI: 10.1016/j.molliq.2016.03.020.
- [26] Yahong, Z., Zhenhua, X., Ximing, W., Li, W., Ai Qin, W. (2012). Adsorption of Congo Red on to Lognocellulose/Montmorillonite Nanocomposite. *J. Wuhan University Technol. Mater.*, 27(5), 931-938. DOI: 10.1007/s11595-012-0576-2.
- [27] Aksu, Z., (2005). Application of biosorption for the removal of organic pollutants: a review. *Process Biochem.*, 40, 997-1026. DOI: 10.1016/j.procbio.2004.04.008.
- [28] Lima, H.K., Tenga, T.T., Ibrahima, M.H., Ahmad, A., Chee, H.T. (2012). Adsorption and Removal of Zinc(II) from Aqueous Solution Using Powdered Fish Bones. *APCBEE Procedia*, 1, 96 - 102. DOI: 10.1016/j.apcb.2012.03.017.
- [29] Mohanty, K., Jha, M., Meikap, B.C., Biswas, M.N. (2006). Biosorption of Cr(IV) from aqueous solutions by Eichhornia crassipes. *Chem. Eng. J.*, 117, 71-77. DOI: 10.1016/j.cej.2005.11.018.
- [30] Mahmoodi, N.M., Kharramfar, S., Najafi, N. (2011). Amine-functionalized silica nanoparticle: Preparation, characterization and anionic dye removal ability. *Desalination*, 279, 61-68. DOI: 10.1016/j.desal.2011.05.059.
- [31] Ho, Y.S., McKay, G. (1999). Pseudo-second order model for sorption processes. *Process Biochem.*, 34, 451-465. DOI: 10.1016/S0032-9592(98)00112-5.
- [32] Yang, L., Zhang, Y., Liu, X., Jiang, X., Zhang, Z., Zhang, T., Zhang, L. (2014). The Investigation of Synergistic and competitive interaction between dye Congo red and Methyl blue on magnetic MnFe<sub>2</sub>O<sub>4</sub>. *Chem. Eng. J.*, 246, 88-96. DOI: 10.1016/j.cej.2014.02.044.
- [33] Khaniabadi, Y.O., Basiri, H., Nourmoradi, H., Mohammadi, M.J., Yari, A.R., Sadeghi, S., Amrane, A. (2017). Adsorption of Congo Red Dye From Aqueous Solutions by Montmorillonite as a Low-cost Adsorbent. *Inter. J. Chem. React. Eng.*, 16(1), 1-11. DOI: 10.15171/EHEM.2017.05.
- [34] Sharma, P.K., Ayub, S., Tripathi, C.N. (2016). Isotherm describing physical adsorption of Cr(VI) from aqueous solution using various agricultural wastes as adsorbent. *Cogent. Eng.*, 3, 1-20. DOI: 10.1080/23311916.2016.1186857.
- [35] Baccar, R., Blangquez, P., Bouzid, J., Feki, M., Sarra, M. (2010). Equilibrium, thermodynamic and kinetic studies on adsorption of commercial dye by activated carbon derived from olive-waste cakes. *Chem. Eng. J.*, 165(2), 457-464. DOI: 10.1016/j.cej.2010.09.033.
- [36] Gao, J.-J., Qin, Y.-B., Zhou, T., Cao, D.-D., Xu, P., Hochstetter, D., Wang, Y.-F. (2013). Adsorption of methylene blue onto activated carbon produced from tea (*Camellia sinensis*

- L.) seed shells: kinetics, equilibrium, and thermodynamics studies. *J.f Zhejiang University: Science B (Biomedicine & Biotechnol.)*, 14(7), 650-658. DOI: 10.1631/jzus.B12a0225.
- [37] Nurhadi, M., Chandren, S., Yuan, L. S., Ho, C. S., Mahlia, T. M. I., Nur, H. (2017). Titania-Loaded Coal Char as Catalyst in Oxidation of Styrene with Aqueous Hydrogen Peroxide. *Int. J. Chem. Reactor Eng.*, 15(1), 1-11. doi: 10.1515/ijcre-2016-0088.
- [38] Geng, L., Wang, Y., Yu, G., Zhu, Y. (2011). Efficient carbon-based solid acid catalysts for the esterification of oleic acid. *Catal. Commun.*, 13(1), 26-30. DOI: 10.1016/j.catcom.2011.06.014.
- [39] Nurhadi, M., Kusumawardani, R., Nur, H. (2018). Negative Effect of Calcination to Catalytic Performance of Coal Char-loaded TiO<sub>2</sub> Catalyst in Styrene Oxidation with Hydrogen Peroxide as Oxidant. *Bull. Chem. React. Eng. Catal.*, 13(1), 113-118. DOI: 10.9767/bcrec.13.1.1171.113-118.
- [40] Nurhadi, M. (2017). Modification of Coal Char-loaded TiO<sub>2</sub> by Sulfonation and Alkylsilylation to Enhance Catalytic Activity in Styrene Oxidation with Hydrogen Peroxide as Oxidant. *Bull. Chem. React. Eng. Catal.*, 12(1), 55-61. DOI: 10.9767/bcrec.12.1.501.55-61.
- [41] Peng, L., Philippaerts, A., Ke, X., Van Noyen, J., De Clippel, F., Van Tendeloo, G., Sels, B.F. (2010). Preparation of sulfonated ordered mesoporous carbon and its use for the esterification of fatty acids. *Catal. Today*, 150(1-2), 140-146. DOI: 10.1016/j.cattod.2009.07.066.
- [42] Qi, Z., Wenqi, G., Chuanxin, X., Dongjiang, Y., Xiaoqing, L., Xiao, Y. Xiaofang, L. (2011). Removal of Neutral Red from aqueous solution by adsorption on spent cottonseed hull substrate. *J. Hazard. Mater.*, 185, 502-506. DOI: 10.1016/j.jhazmat.2010.09.029.

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